



## Exposure to indoor air pollutants during physical activity in fitness centers



C.A. Ramos <sup>a, b, \*</sup>, H.T. Wolterbeek <sup>b</sup>, S.M. Almeida <sup>a</sup>

<sup>a</sup> Centro de Ciências e Tecnologias Nucleares, Instituto Superior Técnico, Universidade de Lisboa, Estrada Nacional 10, Km 139.7, 2695-066 Bobadela LRS, Portugal

<sup>b</sup> Faculty of Applied Sciences, Department of Radiation, Radionuclides and Reactors, Technical University of Delft, Delft, The Netherlands

### ARTICLE INFO

#### Article history:

Received 23 June 2014

Received in revised form

1 August 2014

Accepted 22 August 2014

Available online 2 September 2014

#### Keywords:

Fitness centers

Indoor air quality

Exposure

Chemical pollutants

Physical activity

### ABSTRACT

Physical activity has become a social need among people and it has been clearly proved that exercise is a way to prevent all-cause and cardiovascular-related death, diabetes mellitus and obesity. However, athletes and the common individual can be at risk when they are practicing exercise in polluted environments. In 2012, a monitoring program was undertaken in 11 fitness centers from Lisbon where comfort parameters (temperature and humidity) and indoor air pollutants (PM<sub>10</sub>, PM<sub>2.5</sub>, CO<sub>2</sub>, CO, CH<sub>2</sub>O and VOC) were measured. Three gyms were selected to perform a deeper analysis consisting of longer measurement periods and more parameters, such as particle chemical composition and nanoparticle lung deposition. Measurements were performed during the occupation time, in the studios and in the bodybuilding room, in order to recognize daily patterns and to identify pollutant sources. The pollutants CO<sub>2</sub>, VOC and CH<sub>2</sub>O presented high concentrations exceeding the national limit values, while O<sub>3</sub> and CO did not present concerning levels. Pollutant continuous measurements demonstrated increased levels of particles when the spaces were occupied during classes. Results indicated that it is crucial to optimize the HVAC systems, ventilation rates and occupants behavior in order to reduce the exposure to air pollutants in fitness centers and to potentiate the benefits of sport activities.

© 2014 Elsevier Ltd. All rights reserved.

### 1. Introduction

According to the World Health Organization, physical inactivity was identified as the fourth greatest risk factor for mortality, accounting for 3.2 million deaths per year in the world [1]. Physical activity is an important factor for life quality and frequent practice of exercise, like walking or bicycling, presents great benefits for health [2]. A clean environment is also essential for human health and well-being. In Europe, the environmental aspect with most concern on human health is related to indoor and outdoor air pollution [3,4]. Considering these two aspects and in order to potentiate the benefits of physical activity, people who live in urban areas choose less polluted outdoor spots or fitness centers to avoid air pollution. In fact, athletes and the common individual can be at risk when they are practicing exercise in polluted environments

due the fact that 1) the increase in the minute ventilation increases proportionally the quantity of inhaled pollutants; 2) most of the air is inhaled through the mouth, bypassing the normal nasal mechanisms for filtration of large particles and 3) the increased airflow velocity carries gaseous pollutants deeper into the respiratory tract [5].

Fitness centers present specific characteristics that can affect the indoor air quality (IAQ). Like in other indoor places, IAQ in gymnasiums is affected by building maintenance, building materials and type of ventilation, but what makes these places peculiar are the higher human occupancy and the type of activity developing inside. Fitness centers join all the conditions that promote the increase of CO<sub>2</sub> concentrations because occupants are the dominant source of indoor CO<sub>2</sub> and its production rate depends primarily on the number of people in the room and on their metabolic level [6]. Occupancy also has influence in the PM<sub>10</sub> concentrations [7]. Braniš and colleagues [8] observed a direct relation between the indoor concentrations of coarse PM and the number of children present in a scholar gymnasium. Also in school gyms, Buonanno [9] concluded that the high levels of coarse PM concentrations are produced by students' activity.

\* Corresponding author. Centro de Ciências e Tecnologias Nucleares, Instituto Superior Técnico, Universidade de Lisboa, Estrada Nacional 10, Km 139.7, 2695-066 Bobadela LRS, Portugal.

E-mail address: [carla.ramos@ctn.ist.utl.pt](mailto:carla.ramos@ctn.ist.utl.pt) (C.A. Ramos).

Therefore, there is an indubitable higher exposure to air pollutants in gymnasiums not only due to the characteristics of these places but also due to the changes in the respiratory parameters caused by the physical activity. However, despite the importance of healthy air in sport facilities, IAQ studies have been focused principally on schools [10–18], elderly care centers [19–21]; homes [22] and offices [23]. Comparatively, IAQ monitoring programs carried out in sport facilities are very scarce. Aside from the studies of Buonanno [9] and Braniš [8,24] performed in school gyms, only the exposure in ice rinks are object of study since the 90's due to the high levels of CO, NO<sub>2</sub> and PM that are emitted by the ice resurface vehicles [25–27] and exposure in swimming-pools due to the high concentrations of trihalomethanes [28]. Some other works were conducted in fitness centers but their focus was energy consumption or thermal comfort [29–31].

This paper aims to conduct a comprehensive characterization of IAQ in fitness centers and to identify the principal sources that compromise IAQ. This evaluation will be useful for epidemiological studies and to develop appropriate control strategies not only to minimize the adverse health effects on exercise practitioners, but also to potentiate the benefits of the physical activity.

## 2. Methodology

### 2.1. IAQ assessment in 11 fitness centers

A monitoring program was undertaken in 11 fitness centers from Lisbon where comfort parameters (temperature and humidity) and indoor air pollutants (PM, CO<sub>2</sub>, CO, CH<sub>2</sub>O, VOC and O<sub>3</sub>) were measured. Fig. 1 shows the localization of the selected fitness centers and Table 1 presents their specific characteristics.

Three direct reading apparatus were used: a Lighthouse Handled 3016 to measure PM<sub>5-10</sub>, PM<sub>2.5-5</sub>, PM<sub>1-2.5</sub>, PM<sub>0.5-1</sub> and PM<sub>0.3-0.5</sub>, temperature and relative humidity, a WolfSense to assess CO<sub>2</sub>, CO, VOC and O<sub>3</sub> and a Formaldemeter htV-M to quantify the levels of CH<sub>2</sub>O. All devices were calibrated according to the fabricators specifications.

Measurements were performed during the period of the late afternoon/night, in order to overlap with the more occupied period and with the worst case scenario. In each fitness center, measurements took place in the bodybuilding room (Bb) for 60 min, and in two studios (S1 and S2), during the time of a fitness class (varied between 45 and 60 min). Equipment was positioned at an elevation of 1.20 m from the ground and 1.50 m away from walls in order to avoid the influence on airborne particle dispersion [32,33]. The same air pollutants described previously were measured in the outdoor air. The sampling campaign was performed in October of 2012.

### 2.2. IAQ assessment in 3 selected fitness centers

After the previous analysis, three fitness centers were selected in order to perform a deeper IAQ assessment. In these sport facilities 1) daily continuous measurements of pollutants were performed in different spaces of the gymnasiums in order to recognize daily patterns and identify pollutant sources; 2) particles were sampled and element concentrations were measured; and 3) nanoparticle deposition in lungs were studied. The adopted methodology resulted in three campaigns of six days assessment, in each fitness center, that occurred from October to December 2012.

The selection of the fitness centers (G9, G10 and G11) was made based on the number of daily users, number of fitness classes during the day and type of ventilation. In G9, two studios and the Bb room were selected and monitored during two days in each

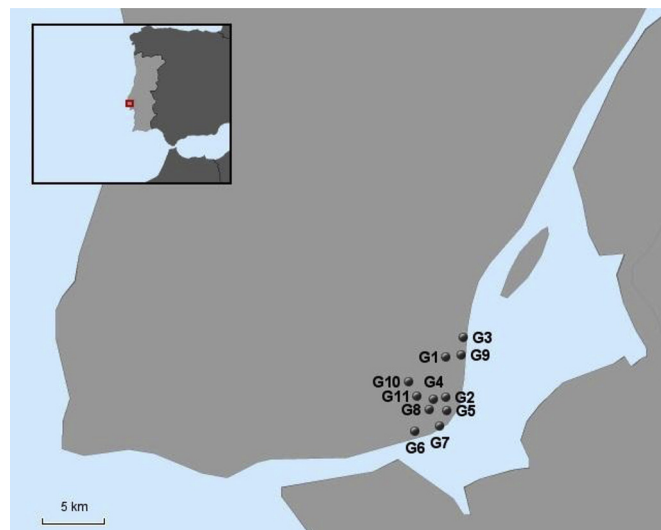


Fig. 1. Localization of the 11 fitness centers (G1-G11) in Lisbon, Portugal.

space. In G10 and G11 only one studio was monitored (since the equipment's noise was incompatible with the classes practiced inside the other studios) resulting in four days of monitoring in the selected studio and two days in the bodybuilding room.

#### 2.2.1. Continuous measurements of gases

The pollutants CO<sub>2</sub>, CO, VOC and O<sub>3</sub> were measured continuously with the equipment WolfSense. Data was registered in the same conditions as in 2.1. In outdoor, parallel measurements of CO<sub>2</sub> and CO were performed with the equipment TSI 7545. Both devices were calibrated according with the fabricator specifications.

#### 2.2.2. Particle sampling and measurement

Within the indoor areas of the selected fitness centers, particles were not only measured continuously with a Lighthouse Handled 3016, but were also sampled for subsequent PM<sub>10</sub> chemical characterization with the medium volume sampler, MVS6 Leckel (flow rate 3.5 m<sup>3</sup>/h). Simultaneously, a Partisol 2000 (flow rate 1 m<sup>3</sup>/h) was used for outdoor PM<sub>10</sub> chemical characterization. Teflon filters with 47 mm were used to collect particles.

When the sampling was conducted in the studios, PM<sub>10</sub> samplers (MVS6 and Partisol 2000) only worked during the occupied time, whereas in the Bb room, these devices worked continuously from the opening until the closure of the fitness centers.

The direct reading apparatus worked continuously, from the opening until the closure of the gymnasiums, and particle concentrations were registered every 60 s.

A correction factor ( $\beta$ ) was applied to the PM indoor concentrations obtained by the Lighthouse Handled 3016. This correction factor was obtained by calculating the ratio between the concentrations obtained by the gravimetric method (considered as the reference method) and the concentrations measured by the Lighthouse 3016 [34,35]. The opening and closing of windows and the number of occupants were registered.

#### 2.2.3. Elemental composition of PM<sub>10</sub>

The filter loads were determined by gravimetry using a 0.1  $\mu$ g-sensitivity balance in a clean laboratory (class 10,000) at a temperature of  $20 \pm 1$  °C and a relative humidity of  $50 \pm 5\%$  [36]. Before being weighted, filters were equilibrated for 24 h in the same room. Filters were weighted before and after sampling and the mass was

**Table 1**  
Main characteristics of the studied gymnasiums.

Gym	Year of construction	Location	Space	Volume (m <sup>3</sup> )	Capacity (person) <sup>a</sup>	Floor type	Wall type	Ventilation system
G1	2009	Urban (residential area)	S1	337	20	Linoleum	Brickwork	Mixed
			S2	448	20			
			Bb	65	40			
G2	1997	Urban (street with intense road traffic)	S1	129	20	Floating floor	Brickwork	Mechanical
			S2	266	30			
			Bb	2442	50			
G3	2003	Urban (residential area)	S1	394	25	Floating floor	Brickwork glass	Mechanical
			S2	394	25			
			Bb	990	40			
G4	1982	Urban (street with intense road traffic)	S1	146	15	Wood	Brickwork	Natural
			S2	136	15			
			Bb	87	10			
G5	2012	Urban (street with intense road traffic)	S1	219	30	Floating floor	Brickwork	Mechanical
			S2	82	15			
			Bb	641	20			
G6	2012	Urban (residential area)	S1	395	35	Linoleum	Brickwork glass	Mechanical
			S2	462	35			
			Bb	1509	50			
G7	1923	Urban (residential area)	S1	387	30	Wood/Linoleum	Brickwork	Mixed
			S2	748	40			
			Bb	866	40			
G8	2012	Urban (street with intense road traffic)	S1	148	20	Linoleum	Brickwork glass	Mechanical
			S2	306	30			
			Bb	1062	50			
G9	2010	Urban (street with intense road traffic)	S1	447	35	Floating floor	Brickwork glass	Mechanical
			S2	788	35			
			Bb	1948	60			
G10	2000	Urban (residential area)	S1	1156	40	Floating floor	Brickwork glass	Natural
			S2	1156	40			
			Bb	540	40			
G11	2005	Urban (inside a city park)	S1	745	35	Linoleum Floating floor	Brickwork glass	Mechanical
			S2	745	35			
			Bb	1843	70			

<sup>a</sup> Maximum capacity. S1 – Studio 1; S2 – Studio 2; Bb – Bodybuilding room.

obtained as the average of three measurements, when observed variations were less than 1%.

The chemical characterization of indoor and outdoor PM<sub>10</sub> samples was performed by Instrumental Neutron Activation Analysis using the k<sub>0</sub> methodology (k<sub>0</sub>-INAA) [37–39].

For k<sub>0</sub>-INAA, half of a filter was rolled up and put into a clean thin foil of aluminum and irradiated for 5 h at a thermal neutron flux of  $1.03 \times 10^{13}$  cm<sup>2</sup>/s in the Portuguese Research Reactor. After irradiation, the sample was removed from the aluminum foil and transferred to a polyethylene container. For each irradiated sample, two gamma spectra were measured during 7 h with a hyperpure germanium detector: one spectrum 2–3 days after the irradiation and the other after 4 weeks [40,41]. The accuracy of the analytical method was evaluated with the certified reference material NIST-1633a, Coal Fly Ash, revealing results with an agreement of  $\pm 12\%$  [42,43]. During the sampling campaign, 6 blank filters were treated the same way as regular samples. All measured species were homogeneously distributed; therefore, concentrations were corrected by subtracting the filter blank contents.

#### 2.2.4. Nanoparticle deposition

Nanoparticles are described as having an increasing surface area with a decreasing particle size for the same amount of mass. Consequently, from the viewpoint of nanoparticle toxicity, the determination of nanoparticle surface area deposited in the human lung is very desirable [44]. Therefore, in this study, a nanoparticle surface area monitor (NSAM) (TSI, Model 3550; Shoreview, MN) was used to measure the lung-deposited surface area of particles which is expressed as square micrometers of lung surface per cubic centimeter of inhaled air ( $\mu\text{m}^2/\text{cm}^3$ ). This deposition corresponds to the tracheobronchial or alveolar regions of the human lung,

according to the International Commission on Radiological Protection deposition model developed by the American Conference of Governmental Industrial Hygienists [45].

This equipment worked continuously in the studios and in the Bb room and was installed at the same conditions as in chapter 2.1 but the data was registered every 10 s.

### 3. Results and discussion

#### 3.1. Part 1: IAQ in 11 fitness centers

Table 2 presents the concentrations obtained in the monitoring program that was undertaken in 11 fitness centers from Lisbon. The average and the range values are presented together with the outdoor air measurements.

The Portuguese legislation, Portaria n.º 353-A/2013 [46], defines indoor air limit values (LV) for the pollutants PM<sub>10</sub>, PM<sub>2.5</sub>, CO<sub>2</sub>, CO, CH<sub>2</sub>O, and VOC, as presented in Table 3. Nevertheless, authors considered it important to include O<sub>3</sub> in this table due to its impact on human health, reactivity with other pollutants, producing sub-micron particles that contribute to total particulate exposures, and indoor sources [47]. In Fig. 2, the results obtained in this work were compared with the LV based on a color scale where a) green corresponds to levels below 75% of the LV, b) yellow relates to concentrations between 75% of the LV and the limit value and c) red corresponds to values higher than the LV.

In the fitness centers G9 and G10, all spaces were classified as green for PM<sub>10</sub> concentrations, representing 18% of the studied group. In G1, G3 and G7, PM<sub>10</sub> concentrations exceeded the LV of 50  $\mu\text{g}/\text{m}^3$ , representing 27% of the evaluated spaces. Except for the fitness centers G4 and G10, PM<sub>10</sub> levels were higher in the outdoor

**Table 2**  
Pollutant concentrations measured in the 11 fitness centers.

Gym		CO (mg/m <sup>3</sup> )	CO <sub>2</sub> (mg/m <sup>3</sup> )	PM <sub>10</sub> (μg/m <sup>3</sup> )	PM <sub>2.5</sub> (μg/m <sup>3</sup> )	PM <sub>1</sub> (μg/m <sup>3</sup> )	VOC (mg/m <sup>3</sup> )	CH <sub>2</sub> O (mg/m <sup>3</sup> )	O <sub>3</sub> (mg/m <sup>3</sup> )	T (°C)	RH (%)
G1	S1	1.5 [1.0–1.7]	2624 [2276–2978]	77 [60–105]	19 [12–31]	8.9 [4.8–16]	–	0.2	0.01 [0–0.02]	21 [19–22]	72 [64–82]
	S2	0.6 [0.3–1.3]	1911 [1511–2682]	54 [41–88]	17 [15–23]	12 [10–15]	–	0.17	0.01 [0–0.01]	21 [21–22]	64 [61–70]
	Bb	1.3 [1.0–1.6]	2542 [2148–2992]	61 [47–74]	17 [14–20]	11 [8.4–13]	–	0.23	0.01 [0–0.01]	22 [22–23]	64 [62–67]
	Out	0.4	861	–	–	–	–	0.19	0.02	18	56
G2	S1	0.89 [0.40–1.40]	1181 [988–1373]	31 [24–39]	10 [10–11]	3.5 [3.04–3.8]	0	0.04	0	17 [17–16]	45 [47–44]
	S2	0.087 [0.00–0.30]	1665 [1564–1860]	47 [34–103]	12 [11–14]	4.4 [3.7–4.8]	0	0.08	0	18 [18–18]	51 [50–52]
	Bb	1.7 [1.6–1.8]	1430 [1363–1557]	33 [29–37]	8.9 [9.2–8.8]	2.8 [2.7–2.8]	0.45 [0.04–0.89]	0.04	0	15 [14–16]	55 [51–55]
	Out	–	–	26 [24–28]	11 [11–11]	3.6 [3.5–3.6]	–	–	–	–	–
G3	S1	0.31 [0.20–0.40]	1789 [987–2299]	101 [45–153]	23 [16–27]	5.2 [4.3–5.9]	1.2 [0.92–1.4]	0.04	0	18 [17–19]	53 [48–57]
	S2	0	1993 [1813–2299]	89 [63–143]	23 [31–19]	5.6 [5.1–6.9]	1.02 [0.99–1.1]	0.04	0	20 [19–20]	53 [53–53]
	Bb	0.78 [0.40–1.2]	1069 [952–1619]	65 [52–76]	20 [18–21]	4.7 [4.6–4.9]	1.15 [0.94–1.44]	0.04	0	16 [16–17]	50 [49–51]
	Out	1.9 [1.8–2.0]	524 [456–597]	49 [42–55]	11 [12–10]	3.4 [3.5–3.2]	0.87 [0.38–1.09]	0.03	0.01 [0.01–0.02]	12 [11–13]	68 [54–76]
G4	S1	2.6 [2.4–2.7]	2431 [2022–2675]	43 [29–67]	8.9 [8.5–9.7]	2.18 [2.1–2.3]	1.9 [1.7–2.3]	0.25	0	15 [14–16]	73 [67–75]
	S2	1.8 [1.7–2.0]	2042 [1122–2986]	35 [26–50]	9.2 [8.1–10]	2.5 [2.2–2.7]	1.5 [1.4–1.7]	0.21	0	18 [16–19]	56 [54–60]
	Bb	2.2 [1.9–2.4]	4418 [3880–5021]	43 [34–52]	11 [9.4–12]	3.7 [4.2–2.7]	1.7 [1.6–1.9]	0.13	0	20 [19–21]	61 [57–65]
	Out	1.2 [1.1–1.5]	896 [859–905]	51 [39–82]	11 [10–12]	3.4 [3.2–3.5]	0.65 [0.61–0.69]	0.06	0.01 [0.01–0.02]	16 [14–17]	45 [41–50]
G5	S1	1.8 [1.5–2.2]	2401 [2077–2640]	49 [47–52]	18 [18–18]	6.8 [6.7–6.9]	2.3 [2.1–2.5]	0.10	0	18 [17–18]	77 [74–81]
	S2	1.8 [1.5–2.1]	4109 [2573–5617]	42 [34–54]	6.6 [7.2–6.1]	16 [15–18]	1.8 [1.3–2.2]	1.4	0.01	19 [18–21]	19 [18–20]
	Bb	2.6 [2.4–2.8]	3139 [2945–3341]	37 [31–44]	11 [10–11]	3.4 [3.3–3.5]	2.2 [1.6–2.5]	1.5	0	17 [16–17]	86 [84–90]
	Out	0.66 [0–2.3]	809 [784–835]	37 [34–44]	18 [16–20]	5.2 [3.8–6.8]	0.93 [0.87–1]	0.05	0.02 [0.01–0.04]	13 [11–15]	60 [53–67]
G6	S1	1.3 [1.2–1.4]	1550 [1363–1720]	17 [9–45]	5.8 [4.9–7.7]	3.5 [3.2–4]	2.03 [1.9–2.1]	0.11	0.02 [0.01–0.03]	17 [18–20]	56 [54–58]
	S2	1.01 [0.60–1.2]	3484 [2336–3932]	44 [21–68]	11 [7.4–13]	4.3 [3.7–4.7]	1.7 [1.5–1.9]	0.08	0	19 [18–20]	76 [69–79]
	Bb	1.2 [1.4–1.1]	1414 [1136–1708]	26 [21–37]	7.3 [7–8.6]	4.5 [4.3–4.8]	2 [1.8–2.2]	0.09	0.01 [0–0.02]	19 [18–20]	56 [54–58]
	Out	1.7 [1–2.1]	942 [832–999]	26 [24–28]	11 [11–11]	3.8 [3.7–4.04]	1.6 [1.5–1.7]	0.08	0.03 [0.01–0.04]	12 [11–12]	52 [47–57]
G7	S1	0	1732 [381–2835]	–	–	–	0.50 [0.06–0.57]	0.01	0.19 [0.07–0.82]	17 [16–17]	51 [61–40]
	S2	0	2751 [2078–3330]	84 [21–105]	11 [4.4–13.7]	3.3 [2.2–3.7]	0.57 [0.34–0.74]	0.01	0	16 [16–16]	68 [59–76]
	Bb	0	1660 [1811–1562]	55 [42–68]	15 [12–17]	5.5 [5.9–4.9]	0.37 [0.33–0.43]	0.01	0.01 [0–0–0.2]	17 [16–18]	53 [51–58]
	Out	0.93 [0.0–2.0]	–	11 [9.4–13]	7 [6.3–7.9]	5.4 [5–5.7]	0.34 [0.31–0.37]	0.01	1.7 [0.18–2.9]	8 [6–10]	44 [38–51]
G8	S1	0.018 [0.0–0.20]	3078 [1173–5964]	50 [19–83]	9.2 [6.3–13]	3 [2.6–4.1]	0	0.01	0	18 [17–19]	80 [64–95]
	S2	0.29 [0.20–0.40]	4234 [3803–4694]	56 [43–67]	10 [9.1–11]	3.2 [2.9–3.4]	0	0.01	0	20 [20–21]	74 [70–79]
	Bb	0	1193 [1045–1381]	29 [23–33]	5.7 [5.1–6.8]	2 [1.6–2.8]	0	0.01	0	17 [16–18]	64 [61–66]
	Out	0	–	7.6 [1.7–21]	3.7 [1.7–9.1]	2.1 [1.03–6.7]	0	0.01	0	10 [10–12]	66 [65–68]
G9	S1	0	1339 [810–1774]	34 [12–49]	7.7 [3.9–14]	4 [2.2–6.5]	2.2 [2–2.5]	0.01	0.01 [0–0.01]	18 [18–19]	73 [71–74]
	S2	0.080 [0.0–0.2]	1266 [860–1735]	17 [4.9–49]	4.1 [1.9–6.9]	2.3 [1.2–3.9]	3.3 [2.9–3.8]	0.02	0.02 [0.01–0.02]	18 [17–19]	71 [67–77]
	Bb	0.10 [0.10–0.10]	2210 [669–3590]	24 [6.8–61]	5.3 [3.2–11]	2.5 [1.5–5.1]	1.9 [1.8–2.2]	0	0.01 [0.01–0.02]	19 [18–20]	70 [66–77]
	Out	0	753 [734–845]	18 [13–23]	4.0 [3.9–4.1]	2.4 [2.4–2.4]	0.65 [0.5–1.2]	0.06	0.06 [0.05–0.09]	22 [22–23]	40 [38–44]
G10	S1	0.28 [0.10–0.50]	1549 [1139–2149]	15 [2.8–25]	12 [10–16]	5.4 [4.7–6.2]	1.8 [1.2–2.3]	0.06	0.02 [0–0.05]	25 [24–25]	43 [41–46]
	S2	0.15 [0.10–0.20]	1277 [984–1482]	3.5 [1.8–8.6]	12 [25–43]	7.3 [1–13]	1.04 [0.95–1.17]	0.06	0.02 [0.01–0.03]	24 [23–25]	44 [40–49]
	Bb	0	1479 [755–2510]	14 [13–18]	14 [13–15.5]	8.1 [7.7–8.8]	1.03 [0.88–1.1]	0.03	0.02 [0.01–0.03]	24 [24–24]	52 [50–54]
	Out	2.3 [1.5–2.9]	899 [748–770]	50 [23–115]	8.8 [6.4–12]	3.8 [3.1–6.7]	2.1 [1.8–2.8]	0.09	0.05 [0.02–0.07]	23 [23–24]	35 [34–36]
G11	S1	0.51 [0.20–0.60]	1116 [673–1652]	79 [57–126]	4 [9–1]	1.5 [3.4–0.7]	1.9 [1.7–2.2]	0.10	0.02 [0.01–0.05]	21 [20–21]	73 [62–85]
	S2	0.53 [0.30–0.80]	1188 [635–1906]	48 [35–71]	1.2 [0.90–2.4]	0.9 [0.74–1.3]	1.8 [1.7–2.02]	0.10	0.02 [0–0.03]	20 [19–21]	76 [74–84]
	Bb	0.68 [0.40–1.10]	1467 [665–2552]	90 [25–71]	7.3 [6.1–8.9]	4.8 [4.1–5.7]	2.5 [2.4–2.5]	0.12	0.01 [0–0.02]	21 [20–21]	67 [66–71]
	Out	1.6 [1.2–2.0]	–	48 [19–108]	15 [12–34]	6.9 [6.4–16]	1.3 [1.2–1.4]	–	0.06 [0.02–0.09]	21 [20–22]	44 [40–47]

Empty spaces were caused by failures in the equipment.

**Table 3**  
Limit values of indoor air pollutants defined by the Portuguese legislation, Portaria n.º 353-A/2013.

Pollutant	Limit value
PM <sub>10</sub>	50 µg/m <sup>3</sup> <sup>(a)</sup>
PM <sub>2.5</sub>	25 µg/m <sup>3</sup> <sup>(a)</sup>
CO <sub>2</sub>	2250 mg/m <sup>3</sup> <sup>(b)</sup>
CO	10 mg/m <sup>3</sup> <sup>(a)</sup>
O <sub>3</sub>	0.2 mg/m <sup>3</sup> <sup>(a)</sup>
CH <sub>2</sub> O	0.1 mg/m <sup>3</sup> <sup>(a)</sup>
VOC	0.6 mg/m <sup>3</sup> <sup>(a)</sup>

<sup>a</sup> Based on the temporal maximum.  
<sup>b</sup> Based on the temporal average.

than in the indoor. For PM<sub>2.5</sub>, a large part of the fitness centers (82%) presented concentrations in the green zone.

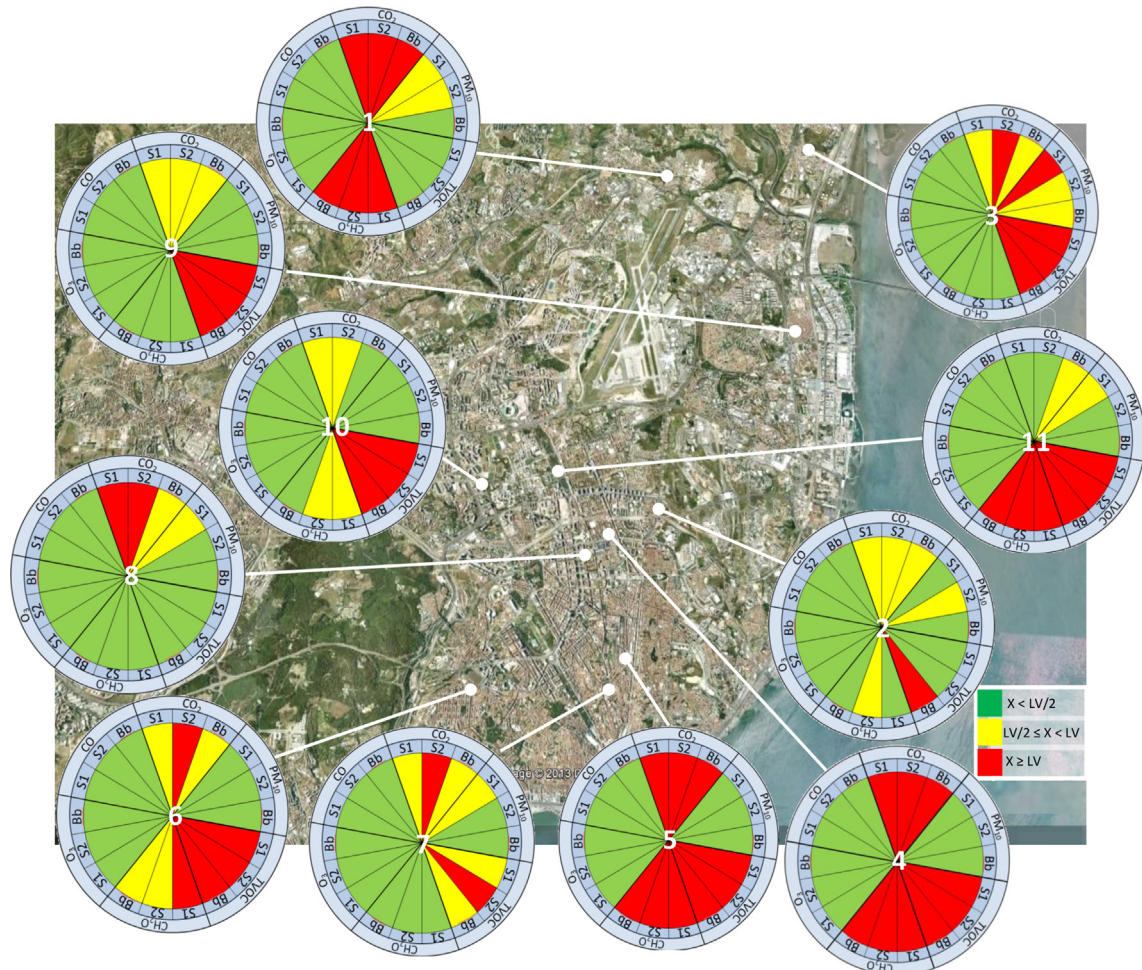
CO levels were always below the LV defined by the Portuguese legislation. CO is principally associated with infiltrations from the outdoors, garages and combustion processes that are principally related to HVAC systems and water heating [48]. Higher concentrations of this pollutant were measured in the indoors of G1, G4, G5, G8 and G9. The observed differences between fitness centers may be explained by the localization of the air intakes of the ventilation systems and by the proximity of the gyms to high traffic roads which contributes to the contamination of the indoor air.

O<sub>3</sub> levels measured in the fitness centers were very low. A maximum concentration of 0.02 mg/m<sup>3</sup> was measured in G6, G9,

G10 and G11. The main O<sub>3</sub> sources in the buildings are the printers [49,50], which are negligible in fitness centers. Outdoors, O<sub>3</sub> occurs as a secondary pollutant, principally as a result from traffic. Therefore, concentrations of this pollutant were always higher outdoor when compared with the indoor environment.

Indoors, the presence of CO<sub>2</sub> is principally associated with occupancy [6]. In 54% of the studied fitness centers, the LV of 2250 mg/m<sup>3</sup> was exceeded in at least one of the spaces. CO<sub>2</sub> average concentration of all spaces was 2000 mg/m<sup>3</sup>. G4 presented the highest average concentrations of CO<sub>2</sub> (4418 mg/m<sup>3</sup>) while the maximum value was reached in G5 (5617 mg/m<sup>3</sup> in studio 2). It is not easy to properly characterize the CO<sub>2</sub> present indoors, since its concentration is a function of the occupation of the site, ventilation rates and metabolic activity of the occupants, with these parameters fluctuating with time [51]. CO<sub>2</sub> levels suggested inefficient ventilation of the studied fitness centers.

Since VOC are emitted by consumer products or structures that exist mainly in the indoor environments, such as carpeting, furniture cleaners, paints, perfumes, lacquers and solvents, the concentrations of VOC are usually found to be higher indoors than outdoors [52]. In our study, exceedances of VOC were registered in 82% of the fitness centers and in 64% of the gymnasiums all the spaces presented concentrations higher than the LV. The highest VOC average concentration was registered in G9 with 3.3 mg/m<sup>3</sup>. CH<sub>2</sub>O is a VOC, but given its importance due to the related health effects, it is usually assessed in an individualized form [53]. However, its indoor sources are also similar to the sources of VOC. In the



**Fig. 2.** Classification of the fitness centers according to the Portuguese legislation for IAQ (Portaria n.º 353-A/2013). S1 and S2 – studios; Bb – bodybuilding room.

majority of the cases, the indoor concentrations were higher than outdoors, except the cases of G7, G8, G9 and G10. The highest CH<sub>2</sub>O concentrations were found in G4 studios (0.25 mg/m<sup>3</sup> in S1 and 0.21 mg/m<sup>3</sup> in S2) together with high concentrations of VOC that may be originated by the presence of alcohol base hand disinfectant distributed throughout this gymnasium.

Some fitness centers presented high values of some pollutants related to their design and construction. Table 2 shows that G5 presented high levels for CO (2.6 mg/m<sup>3</sup> in Bb) and furthermore elevated values for CH<sub>2</sub>O (1.4 mg/m<sup>3</sup> in S2 and 1.5 mg/m<sup>3</sup> in Bb) and VOC (2.3 mg/m<sup>3</sup> in S1 and 2.2 mg/m<sup>3</sup> in Bb). The highest VOC and CH<sub>2</sub>O concentrations registered in this recently open (2012) fitness center are probably associated with emissions from the new furniture, material and equipment: VOC concentrations analyzed in new apartments demonstrate a decreasing tendency in indoor VOC concentrations over the 24 month follow-up period [54]. Moreover, G5 is located on the ground floor level of a major building, so its elevated CO levels may have resulted from the inappropriate location of the air admissions of the HVAC system, which are placed near the road and close to the pavement.

### 3.2. Ventilation rates

Ventilation rates were calculated using the build-up method developed by Hanninen [55] which is based on the use of CO<sub>2</sub> as a tracer gas. CO<sub>2</sub> represents an advantage comparing with other tracers since it is emitted by occupants and it is inert. This method is based on the curve fit of CO<sub>2</sub> concentrations and requires inputs of: the indoor and outdoor CO<sub>2</sub> concentrations, the number of occupants and the volume of the space [56].

Air exchange rates and ventilation rates were calculated for all the fitness centers. However, these parameters were only estimated for studios because the Bb rooms did not present the required constant number of occupants necessary to run the build-up method.

Table 4 presents the air exchange rates (AER) and the ventilation rates (VR) and shows that AER varied between 1.4 h<sup>-1</sup> and 4.4 h<sup>-1</sup> and VR ranged between 8.9 and 51.5 lps/person. Since the Portuguese legislation [46] defines VR between 176 and 353 lps/person for fitness centers, the results indicated that no fitness center meet the Portuguese legislation criteria. According to the main national standards in Europe (but despite the lack of unanimity), the AER of 0.5 h<sup>-1</sup> is defined as a threshold below which associations with poor IAQ may occur [57]. In this study, all fitness centers presented a higher AER.

### 3.3. Part 2: IAQ assessment in three fitness centers

Three fitness centers (G9, G10 and G11) were selected in order to perform a deeper IAQ assessment considering longer measurement periods and more parameters.

**Table 4**  
Air exchange rates (h<sup>-1</sup>) and ventilation rates (lps) in the 11 fitness centers.

Gym	Air exchange rate (h <sup>-1</sup> )	Ventilation rate (lps)
G1	1.6	8.9
G2	2.1	11.4
G3	3.1	43.1
G4	1.4	10.2
G5	2.3	14.0
G6	3.5	15.3
G7	–	–
G8	–	–
G9	4.4	29.3
G10	1.6	46.7
G11	2.3	51.5

#### 3.3.1. Continuous measurements of gases

Fig. 3 presents the CO<sub>2</sub> concentrations measured in the three fitness centers. A similar trend was observed in all gyms which was characterized by an increase of CO<sub>2</sub> levels in the studios during the occupied period. However, results showed that CO<sub>2</sub> concentrations were influenced not only by the number of people inside the room but also by their metabolic activity during the fitness classes. Fig. 4 shows the CO<sub>2</sub> growth curve for the same room but in two different fitness classes: Yoga (mind class) and Body Attack (cardio class). Besides the greater number of occupants in the Yoga class (24 people in Yoga versus 20 people in Body Attack), CO<sub>2</sub> concentrations were significantly lower than in Body Attack class. The average CO<sub>2</sub> concentration was 959 mg/m<sup>3</sup> for Yoga and 1774 mg/m<sup>3</sup> for Body Attack. Additionally, the slope of the CO<sub>2</sub> build-up phase in the Body Attack class was higher reflecting a quick growth in the production of this pollutant. In Yoga class, this increase was not observed.

Inside the bodybuilding rooms, CO<sub>2</sub> concentrations also reflected the degree of occupancy. In G11, the highest CO<sub>2</sub> levels were associated with the cycling classes which occurred inside the bodybuilding room behind a folding screen.

Despite not causing toxicity to humans at the registered concentrations [58], CO<sub>2</sub> is a good indicator of IAQ and can influence the human perception of the spaces. Moreover, the performance of people is affected by the concentrations of this gas. Previous studies showed that changes in CO<sub>2</sub> concentrations were associated with statistically significant and meaningful reductions in decision-making performance [59].

Fig. 5 shows that CO concentrations in the three fitness centers were below the LV (10 mg/m<sup>3</sup>) defined by the Portuguese legislation [46]. In G9, CO concentration increased in the late afternoon/night which is the period with more entrances and exits in the gymnasium. The air intake for this fitness center is located near the garage, and this fact can explain the highest contamination of indoor air by the traffic during this period. In G10, CO concentrations were higher during the morning due to the presence of trucks, which unload material for an annex building. In G11, the rises in CO levels were also traffic-related and enhanced by the surroundings, since the gymnasium building was walled by other buildings with a height greater than eight floors causing a canyon effect which lead to a lack of pollutant dispersion [60]. At the registered concentrations, CO does not present harmful health effects to humans, although this pollutant can connect with hemoglobin, replacing the O<sub>2</sub> which in turns reaches the tissues in smaller concentrations [61].

In fitness center G9, the VOC concentrations exceeded the limit value of 0.6 mg/m<sup>3</sup> most of the time (Fig. 6). Fig. 7 shows that cleaning procedures highly contributed for the increase of VOC concentrations.

#### 3.3.2. Levels of particulate matter

Fig. 8 presents the temporal distribution of PM, measured in five ranges (PM<sub>0.3-0.5</sub>, PM<sub>0.5-1</sub>, PM<sub>1-2.5</sub>, PM<sub>2.5-5</sub> and PM<sub>5-10</sub>) in the selected fitness centers. In G9 and G10, the PM<sub>10</sub> and PM<sub>2.5</sub> Portuguese LV of 50 µg/m<sup>3</sup> and 25 µg/m<sup>3</sup>, respectively, were exceeded.

The maximum concentrations in G9 for PM<sub>10</sub> and PM<sub>2.5</sub> were measured in studio 1 (109 µg/m<sup>3</sup> and 30 µg/m<sup>3</sup>, respectively). In G10, the maximum PM<sub>10</sub> concentrations were 157 µg/m<sup>3</sup> in the studio and 190 µg/m<sup>3</sup> in the bodybuilding room. The maximal PM<sub>2.5</sub> value measured in this fitness center was 37.4 µg/m<sup>3</sup> in studio 1.

Results showed that, in the studios, the highest PM concentrations were coincident with the period of fitness classes, revealing a relation between PM concentration and the resuspension of dust caused by the practitioners of physical activity. In scholar gyms,

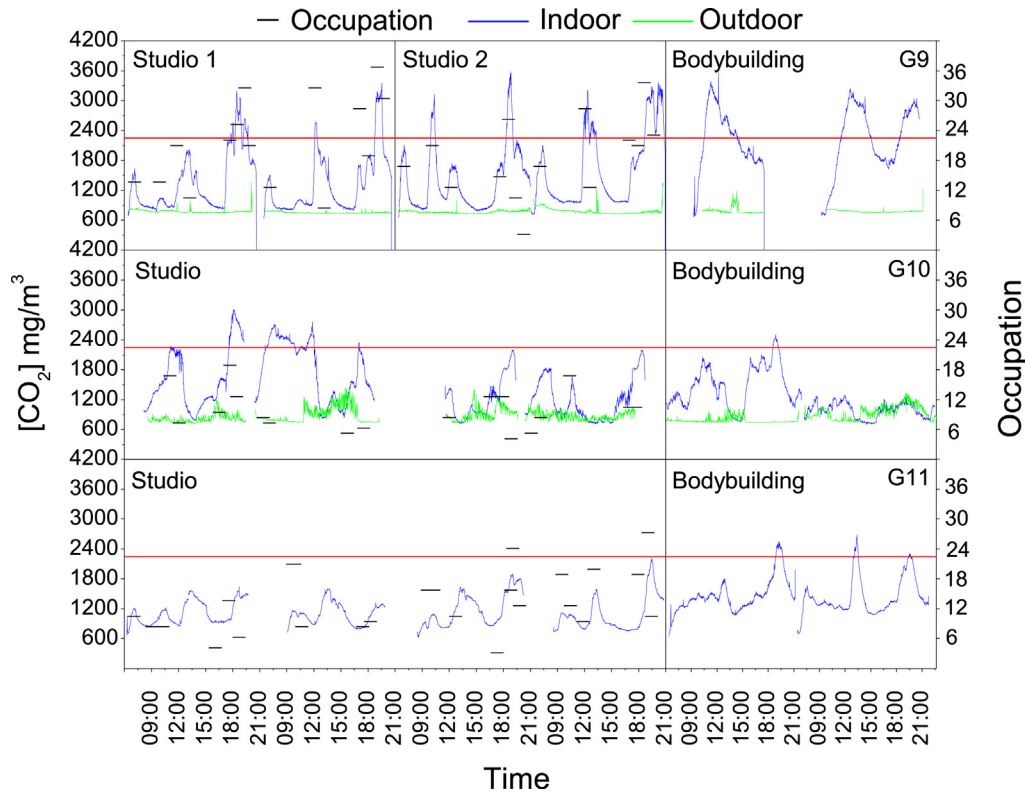


Fig. 3. Temporal variation of CO<sub>2</sub> concentration in the 3 fitness centers (values in mg/m<sup>3</sup>) and human occupation inside the sites. The horizontal line corresponds to the CO<sub>2</sub> LV defined by the Portuguese legislation.

previous studies showed that dust resuspension influenced by students' activity is the major source of coarse particles [8,9].

The highest concentrations measured in the studios of G10 occurred principally during the cleaning operations performed during the afternoon (approximately at 14:00). Cleaning operations have already been identified as one important source for indoor

particle resuspension. Corsi [62] showed that the resuspension caused by vacuum cleaning can increase PM<sub>10</sub> concentrations more than 17 µg/m<sup>3</sup> above the average concentration. Concentrations in the G10 studio increased 8 times in the first day of sampling and 6.5 times in the third day of sampling when compared with the average PM<sub>10</sub> concentrations in the space.

G10 is the only fitness center that opens the windows to ventilate the spaces and this fact was reflected in its highest levels of coarse particles. This gym is placed inside a city park where natural sources of particles, such as soil and pollens, are dominant and contribute principally for the coarse fraction. Canha presented the same conclusion between coarse fraction, natural ventilation and grove vicinity [63].

Fig. 9 presents the comparison between the indoor and outdoor PM<sub>10</sub> total mass concentrations measured by gravimetry. While G9 and G11 presented significantly higher outdoor PM<sub>10</sub> concentrations, in G10 the ratio between indoor and outdoor concentrations was closed to 1 or even higher than 1. These results can be explained by the fact that in G9 and G11 the coarser outdoor particles are retained in the filters, presented in the Air Treatment Units from both buildings, whereas in G10 outdoor air enters in the spaces by the windows without any filtration.

Table 5 shows the indoor and outdoor concentrations for the chemical elements As, Co, Cr, Fe, K, La, Na, Sb, Sc and Zn measured in the PM<sub>10</sub> filters. The outdoor concentrations of these elements were significantly higher than the indoor, except for G10 where significant differences were not observed. This gym has natural ventilation and, consequently, higher contributions of the outdoor elements generated by traffic (As, Sb, Zn), soil (Co, Fe, La and Sc) and sea (Na) were registered [64–67]. Results showed that, besides the higher outdoor As, Co, Cr, Fe, K and Zn concentrations in G9 and G11, the air filtration by their Air Treatment Units allowed

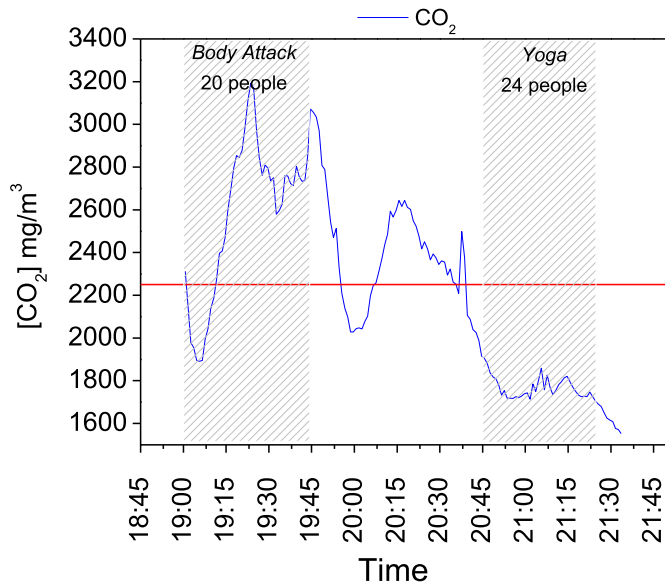


Fig. 4. Growth curve of CO<sub>2</sub> in fitness classes associated with different metabolic rates (values in mg/m<sup>3</sup>). Shading represents the duration of the classes and the horizontal line corresponds to the CO<sub>2</sub> LV defined by the Portuguese legislation.

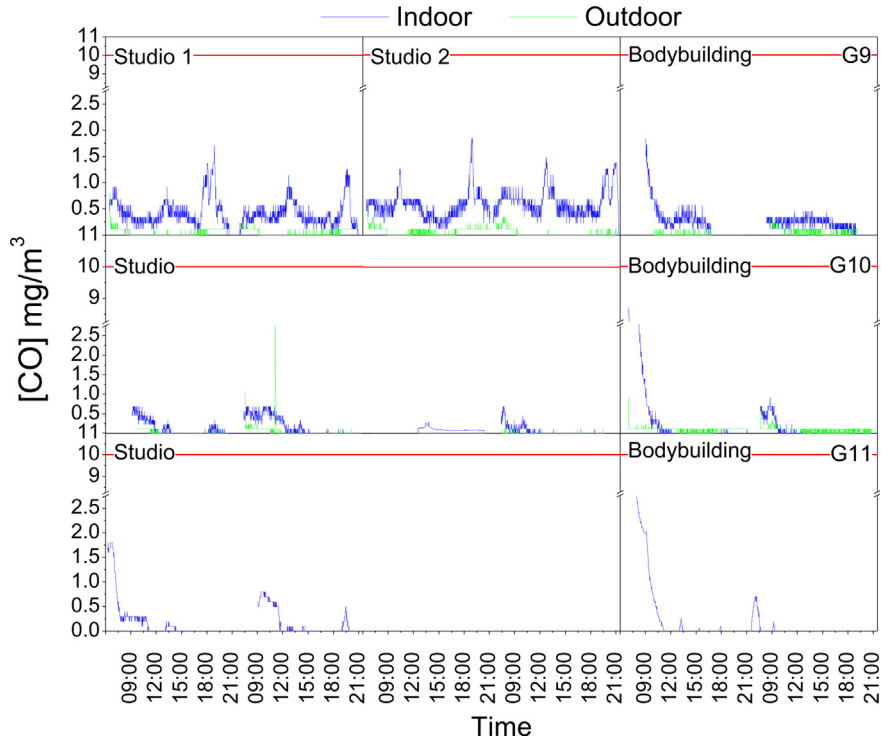


Fig. 5. Temporal variation of CO concentration in the 3 fitness centers (values in  $\text{mg}/\text{m}^3$ ). The horizontal line corresponds to the CO LV defined by the Portuguese legislation.

the retention of particles and, therefore, the capture of these elements.

The crustal enrichment factor method has been used as an attempt to evaluate the strength of the crustal and non-crustal origin of the elements. Enrichment factors, using Fe as a crustal reference element ( $EF_{\text{Fe}}$ ), were calculated based on equation (1) and using soil composition [68]:

$$EF_{\text{Fe}} = \frac{\left(\frac{[x]}{[\text{Fe}]}\right)_{\text{PM}}}{\left(\frac{[x]}{[\text{Fe}]}\right)_{\text{soil}}} \quad (1)$$

Elements with  $EF_{\text{Fe}}$  values that approach unity can be considered predominantly from soil, whereas if the evaluated element has

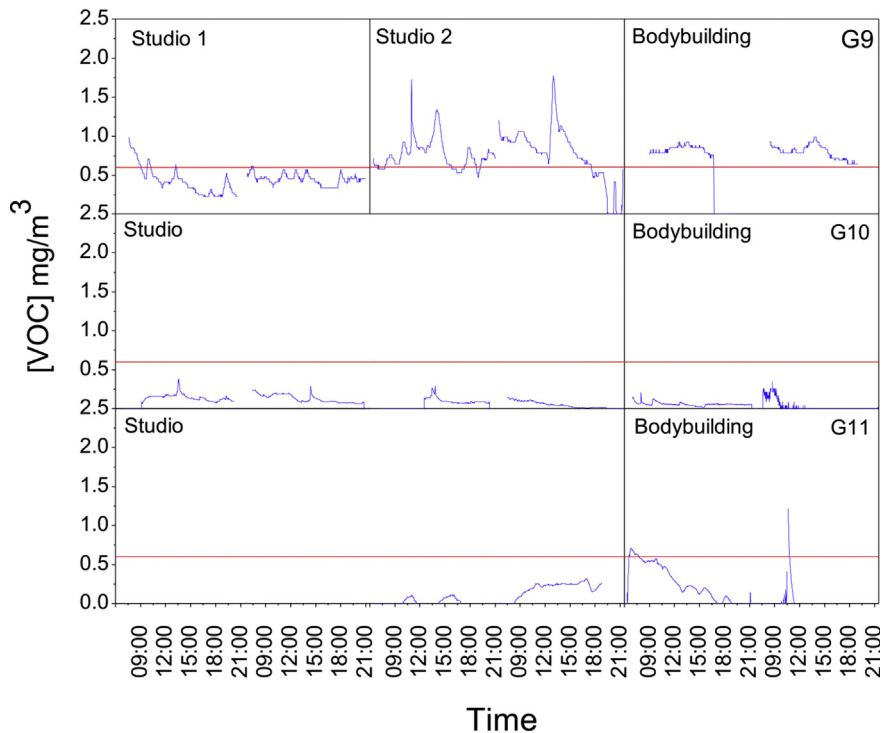
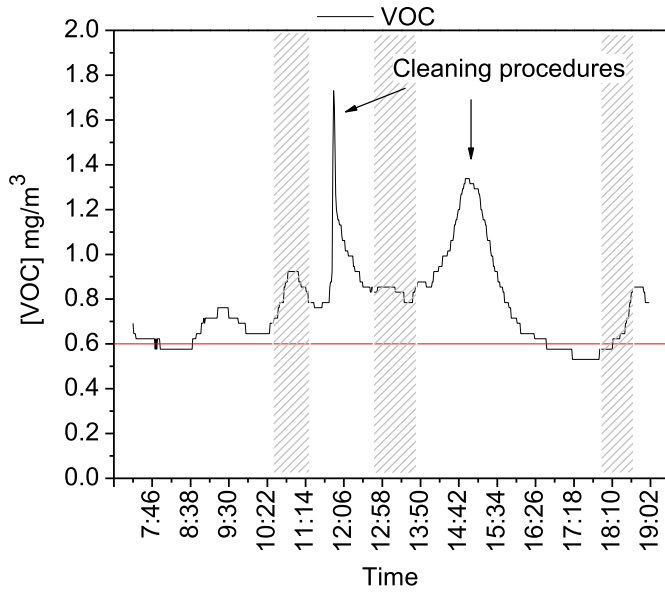
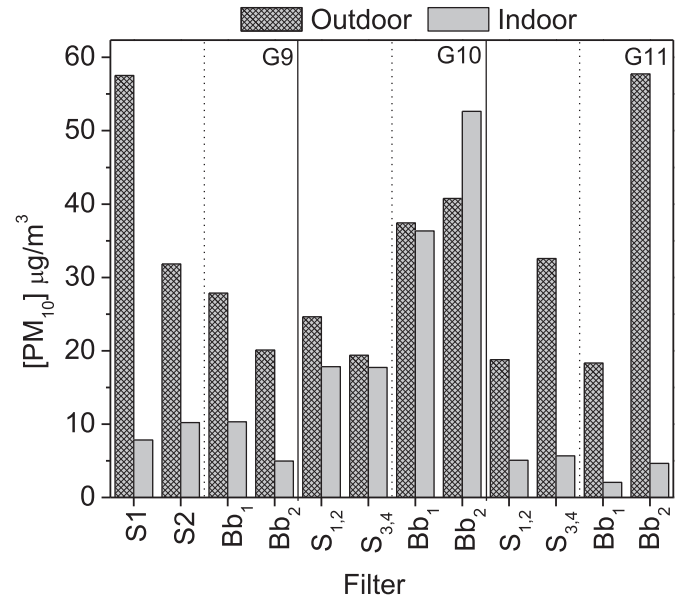


Fig. 6. Temporal variation of VOC concentration in the 3 fitness centers (values in  $\text{mg}/\text{m}^3$ ). The horizontal line corresponds to the VOC LV defined by the Portuguese legislation.





**Fig. 7.** Temporal variation of VOC concentration for a selected period in G9 (values in  $\text{mg}/\text{m}^3$ ). Shading represents the duration of the classes and the horizontal line corresponds to the VOC LV defined by the Portuguese legislation.

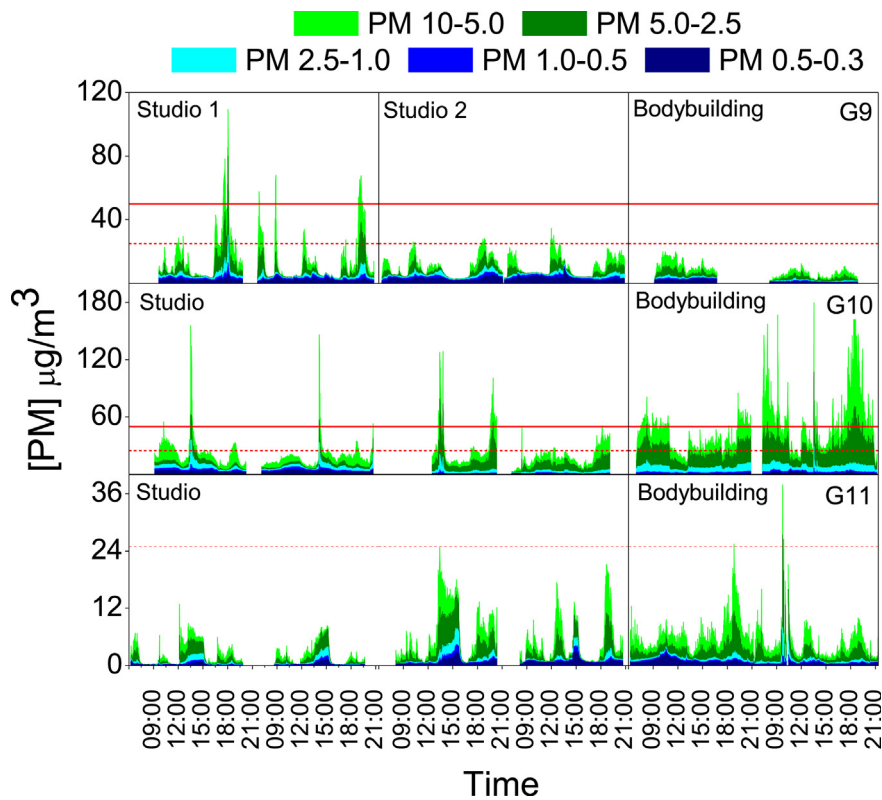


**Fig. 9.**  $\text{PM}_{10}$  concentrations measured indoor and outdoor of the fitness centers (values in  $\mu\text{g}/\text{m}^3$ ). (S1 – Studio 1; S2 – Studio 2; S<sub>1,2</sub> – First and second day of sampling; S<sub>3,4</sub> – third and fourth day of sampling; Bb<sub>1</sub> – First day of sampling in the bodybuilding; Bb<sub>2</sub> – Second day of sampling in the bodybuilding).

EF values higher than 10, its provenance is asserted mainly to local, regional and/or long transportation phenomena from other natural and/or anthropogenic sources [69]. The  $\text{EF}_{\text{Fe}}$ , presented in Fig. 10, indicate that, both in indoor and outdoor, the elements Sc, La, Co, K, Fe and Cr were associated with soil emissions ( $\text{EF}_{\text{Fe}} < 10$ ) while As, Sb and Zn were related to anthropogenic emissions ( $\text{EF}_{\text{Fe}} > 10$ ).

3.3.3. Nanoparticle lung deposition

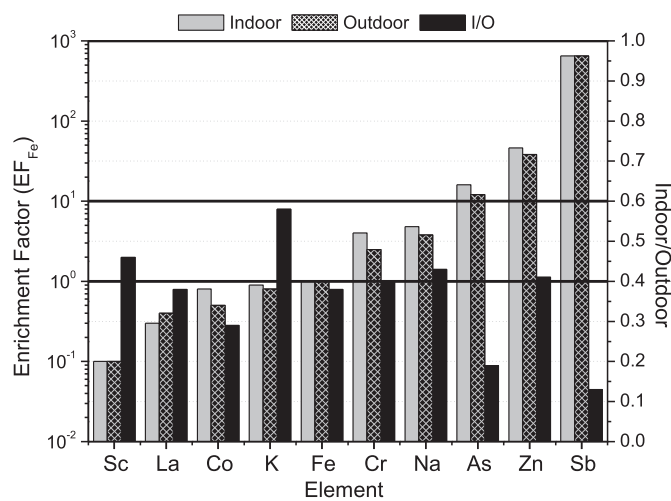
The estimated total deposited alveolar area and the total deposited surface area were calculated for a lung surface area of  $80 \text{ m}^2$ , which is the defined area for an adult. Table 6 shows that the minimum value for the deposited alveolar area was reached in G11



**Fig. 8.** Temporal variation of PM concentration in the 3 fitness centers (values in  $\mu\text{g}/\text{m}^3$ ). The horizontal lines correspond to the  $\text{PM}_{10}$  LV defined by the Portuguese legislation ( $150 \mu\text{g}/\text{m}^3$ ) and to the WHO guidelines to  $\text{PM}_{10}$  ( $100 \mu\text{g}/\text{m}^3$ ) and  $\text{PM}_{2.5}$  ( $50 \mu\text{g}/\text{m}^3$ ).

**Table 5**  
Indoor and outdoor average element concentrations in the fitness centers G9, G10 and G11 (values in ng/m<sup>3</sup>).

	G9		G10		G11		Total	
	I	O	I	O	I	O	I	O
As	0.068 ± 0.032	0.59 ± 0.59	0.25 ± 0.032	0.29 ± 0.041	0.055 ± 0.055	0.41 ± 0.31	0.10 ± 0.088	0.48 ± 0.42
Co	0.06 ± 0.01	0.30 ± 0.22	0.11 ± 0.029	0.14 ± 0.045	0.057 ± 0.049	0.14 ± 0.040	0.076 ± 0.037	0.21 ± 0.16
Cr	1.40 ± 0.86	5.4 ± 3.9	2.47 ± 0.56	3.0 ± 2.1	0.57 ± 0.43	5.5 ± 2.5	1.60 ± 0.98	4.8 ± 2.8
Fe	161 ± 117	1250 ± 1593	375 ± 162	530 ± 350	56 ± 53	771 ± 570	215 ± 190	850 ± 954
K	55 ± 34	480 ± 290	190 ± 22	150 ± 38	51 ± 19	242 ± 257	103 ± 73	291 ± 250
La	0.071 ± 0.0077	0.16 ± 0.14	0.10 ± 0.045	0.16 ± 0.063	0.011 ± 0.010	0.12 ± 0.09	0.064 ± 0.050	0.10 ± 0.09
Na	202 ± 94	640 ± 375	1350 ± 930	1650 ± 760	74 ± 70	691 ± 167	542 ± 770	1020 ± 682
Sb	0.62 ± 0.47	2.61 ± 2.33	0.74 ± 0.16	2.9 ± 1.4	0.19 ± 0.12	2.2 ± 1.7	0.46 ± 0.36	2.5 ± 1.8
Sc	0.0087 ± 0.071	0.014 ± 0.013	0.015 ± 0.0072	0.023 ± 0.0085	<dl	0.0073 ± 0.0099	0.01 ± 0.008	0.02 ± 0.01
Zn	9.7 ± 2.7	74 ± 54	19 ± 11	27 ± 19	5.8 ± 4.7	29 ± 20	12.0 ± 9.0	46 ± 42



**Fig. 10.** Enrichment factor using Fe as a reference element and Mason and Moore (1982) soil composition and ratio indoor/outdoor.

with 13  $\mu\text{m}^2/\text{cm}^3$  and the maximum was registered in G10 with 39  $\mu\text{m}^2/\text{cm}^3$ . As these measurements were performed for the first time in fitness centers, levels were compared with studies performed in other indoor environments. In schools, Buonanno [70] registered higher alveolar area levels deposits which ranged between 35  $\mu\text{m}^2/\text{cm}^3$  and 150  $\mu\text{m}^2/\text{cm}^3$ . In elderly care centers, Almeida-Silva [44] found values between 10  $\mu\text{m}^2/\text{cm}^3$  and 46  $\mu\text{m}^2/\text{cm}^3$  and in houses, Gomes [71] recorded an average value of 29 ± 1.0  $\mu\text{m}^2/\text{cm}^3$  and Ntziachristos [72] registered an average value of 45 ± 26  $\mu\text{m}^2/\text{cm}^3$ .

#### 4. Conclusions

This paper conducted a comprehensive characterization of a vast array of indoor pollutants in 11 fitness centers and identified sources that compromise IAQ.

The high CO<sub>2</sub> levels registered within this study and the calculated ventilation rates indicated that, in general, the fitness centers

**Table 6**  
Average deposited area and total deposited surface area in the fitness centers G9, G10 and G11.

Fitness center	Average deposited area ( $\mu\text{m}^2/\text{cm}^3$ )	Total deposited surface area ( $\mu\text{m}^2$ )
G9	28.61 ± 25.40	1.93 × 10 <sup>7</sup>
G10	39.17 ± 15.95	2.37 × 10 <sup>7</sup>
G11	13.47 ± 6.12	7.99 × 10 <sup>8</sup>

have inefficient ventilation, considering the type of activity that is preconized indoors. This fact influences the human perception of the space and gives the feeling of discomfort during the practice of sports. Taking into account that VOC spikes were observed during cleaning activities and that cleaning products are recognized as risk factors for respiratory health, low emitting agents and “green” practices should be adopted. The levels of particles were highly influenced by the intense indoor activities and by the type of ventilation. Results showed that the location of the air intakes and the efficiency of the air filtration are essential for the maintenance of a good IAQ.

Taking into account the unique characteristics of the fitness centers – intense indoor activities, large number of people who are more susceptible to air pollutants during exercise, insufficient ventilation and relatively small room sizes – there is a need to better assess the exposure and inhaled doses by gyms practitioners in order to minimize adverse health effects and to potentiate the benefits of the physical activity.

#### Acknowledgments

The authors gratefully acknowledge all gyms that collaborated in this study and EFACEC (Eng.º Amádis Santos), Aveiro University (Doutora Célia Alves and Doutora Teresa Nunes) and ESTeSL for supplying part of the equipment. C.A. Ramos acknowledges Fundação para a Ciência e Tecnologia (FCT) for the PhD Grant – SFRH/BD/79277/2011.

#### References

- [1] WHO. Physical inactivity: a global public health problem; 2012. [http://www.who.int/topics/physical\\_activity/en/](http://www.who.int/topics/physical_activity/en/).
- [2] Warburton D, Nicol C, Bredin S. Health benefits of physical activity: the evidence. *CMAJ* 2006;174(6):801–9. <http://dx.doi.org/10.1503/cmaj.051351>.
- [3] EEA. Environment and health; 2011. <http://www.eea.europa.eu/themes/human.intro>.
- [4] Almeida SM, Silva AV, Sarmento S. Effects of exposure to particles and ozone on hospital admissions for cardiorespiratory diseases in Setúbal, Portugal. *J Toxicol Environ Health A* 2014;77(14–16):837–48. <http://dx.doi.org/10.1080/15287394.2014.887399>.
- [5] Carlisle A, Sharp N. Exercise and outdoor ambient air pollution. *Br J Sports Med* 2001;35:214–22. <http://dx.doi.org/10.1136/bjism.35.4.214>.
- [6] Apte MG, Fisk W, Daisey JM. Associations between indoor CO<sub>2</sub> concentrations and sick building syndrome symptoms in U.S. office buildings: an analysis of the 1994–1996 BASE study data. *Indoor Air* 2000;10:246–57. <http://dx.doi.org/10.1034/j.1600-0668.2000.010004246.x>.
- [7] Ferro A, Kopperund R, Hildemann L. Source strengths for indoor human activities that resuspend particulate matter. *Environ Sci Technol* 2004;38:1759–64. <http://dx.doi.org/10.1021/es0263893>.
- [8] Braniš M, Safránek J. Characterization of coarse particulate matter in school gyms. *Environ Res* 2011;111:485–91. <http://dx.doi.org/10.1016/j.envres.2011.03.010>.
- [9] Buonanno G, Fuoco F, Marini S, Stabile L. Particle resuspension in school gyms during physical activities. *Aerosol Air Qual Res* 2013;12:803–13. <http://dx.doi.org/10.4209/aaqr.2011.11.0209>.

- [10] Pegas PN, Evtyugina MG, Alves CA, Nunes T, Cerqueira M, Franchi M, et al. Outdoor/indoor air quality in primary schools in Lisbon: a preliminary study. *Quim Nova* 2010;33:1145–9.
- [11] Canha N, Freitas MC, Almeida SM, Almeida M, Ribeiro M, Galinha C, et al. Indoor school environment: easy and low cost to assess inorganic pollutants. *J Radioanal Nucl Chem* 2010;286(2):495–500. <http://dx.doi.org/10.1007/s10967-010-0781-4>.
- [12] Canha N, Almeida M, Freitas MC, Almeida SM. Seasonal variation of total particulate matter and children respiratory diseases at Lisbon Basic schools using passive methods. *Proc Environ Sci* 2011;4:170–83. <http://dx.doi.org/10.1016/j.proenv.2011.03.021>.
- [13] Almeida SM, Canha N, Silva A, Freitas MC, Pegas P, Alves C, et al. Children exposure to air particulate matter in indoor of Lisbon primary schools. *Atmos Environ* 2011;45:7594–9. <http://dx.doi.org/10.1016/j.atmosenv.2010.11.052>.
- [14] Pegas PN, Alves CA, Evtyugina MG, Nunes T, Cerqueira M, Franchi M, et al. Seasonal evaluation of outdoor/indoor air quality in primary schools in Lisbon. *J Environ Monit* 2011;13:657–67. <http://dx.doi.org/10.1039/C0EM00472C>.
- [15] Canha N, Martinho M, Almeida-Silva M, Freitas MC, Almeida SM, Pegas P, et al. Indoor air quality in primary schools. *Int J Environ Pollut* 2012;50(1/2/3/4):396–410. <http://dx.doi.org/10.1504/IJEP.2012.051210>.
- [16] Canha N, Almeida-Silva M, Freitas MC, Almeida SM. Lichens as biomonitors at indoor environments of primary schools. *J Radioanal Nucl Chem* 2012;291(1):123–8. <http://dx.doi.org/10.1007/s10967-011-1259-8>.
- [17] Canha N, Almeida SM, Freitas MC, Trancoso M, Sousa A, Mouro F, et al. Particulate matter in indoor environments of urban and rural primary schools, by a passive sampling methodology. *Atmos Environ* 2014;83:21–34. <http://dx.doi.org/10.1016/j.atmosenv.2013.10.061>.
- [18] Canha N, Almeida SM, Freitas MC, Wolterbeek HT, Cardoso J, Pio C. Impact of wood burning on window PM<sub>2.5</sub> in a primary school in rural Portugal. *Atmos Environ* 2014;94:663–70. <http://dx.doi.org/10.1016/j.atmosenv.2014.05.080>.
- [19] Almeida-Silva M, Wolterbeek HT, Almeida SM. Elderly exposure to indoor air pollutants. *Atmos Environ* 2014;85:54–63. <http://dx.doi.org/10.1016/j.atmosenv.2013.11.061>.
- [20] Viegas C, Almeida-Silva M, Gomes AQ, Wolterbeek HT, Almeida SM. Fungal contamination assessment in Portuguese elderly care centers. *J Toxicol Environ Health A* 2014;77(1–3):14–23. <http://dx.doi.org/10.1080/15287394.2014.861336>.
- [21] Almeida-Silva M, Almeida SM, Wolterbeek HT. Multi-elemental characterization of indoor aerosols in Elderly Care Centers. *J Radioanal Nucl Chem* 2014;300:679–84. <http://dx.doi.org/10.1007/s10967-014-2997-1>.
- [22] Osman LM, Douglas JG, Garden C, Reglitz K, Lyon J, Gordon S, et al. Indoor air quality in homes of patients with chronic obstructive pulmonary disease. *Am J Resp Crit Care* 2007;176:465–72. <http://dx.doi.org/10.1164/rccm.200605-5890C>.
- [23] Bluyssen PM, Fernandes ED, Groes L, Clausen G, Fanger PO, Valbjorn O, et al. European indoor air quality audit project in 56 office buildings. *Indoor Air* 1996;6:221–38. <http://dx.doi.org/10.1111/j.1600-0668.1996.00020.x>.
- [24] Braniš M, Safránek J, Hytychova A. Indoor and outdoor sources of size-resolved mass concentration of particulate matter in a school gym – implications for exposure of exercising children. *Environ Sci Pollut R* 2011;18:598–609. <http://dx.doi.org/10.1007/s11356-010-0405-0>.
- [25] Lee K, Yanagisawa Y, Spengler JD, Nakai S. Carbon monoxide and nitrogen dioxide exposures in indoor ice skating rinks. *J Sports Sci* 1994;12:279–83. <http://dx.doi.org/10.1080/02640419408732173>.
- [26] Yang C, Demokritou P, Chen Q, Spengler J, Parsons A. Ventilation and air quality in indoor ice skating arenas. *ASHRAE Trans* 2000;106:338–46.
- [27] Pelham T, Holt L, Moss M. Exposure to carbon monoxide and nitrogen dioxide in enclosed ice arenas. *Occup Environ Med* 2002;59(4):224–33. <http://dx.doi.org/10.1136/oem.59.4.224>.
- [28] Dyck R, Sadiq R, Rodriguez MJ, Simard S, Tardif R. Trihalomethane exposures in indoor swimming pools: a level III fugacity model. *Water Res* 2011;45:5084–98. <http://dx.doi.org/10.1016/j.watres.2011.07.005>.
- [29] Lam JC, Chan ALS. CFD analysis and energy simulation of a gymnasium. *Built Environ* 2001;36(3):351–8. [http://dx.doi.org/10.1016/S0360-1323\(00\)00014-7](http://dx.doi.org/10.1016/S0360-1323(00)00014-7).
- [30] Beusker E, Stoy C, Pollalis SN. Estimation model and benchmarks for heating energy consumption of schools and sport facilities in Germany. *Built Environ* 2012;49:324–35. <http://dx.doi.org/10.1016/j.buildenv.2011.08.006>.
- [31] Revel GM, Arnesano M. Perception of the thermal environment in sports facilities through subjective approach. *Built Environ* 2014;77:12–9. <http://dx.doi.org/10.1016/j.buildenv.2014.03.017>.
- [32] Jin H, He C, Lu L, Fan J. Numerical investigation of the wall effect on airborne particle dispersion in a test chamber. *Aerosol Air Qual Res* 2013;13:786–94. <http://dx.doi.org/10.4209/aaqr.2012.04.0106>.
- [33] Holmberg S, Li Y. Modeling of the indoor environment – particle dispersion and deposition. *Indoor Air* 1998;8:113–22. <http://dx.doi.org/10.1111/j.1600-0668.1998.t01-2-00006.x>.
- [34] McNamara M, Noonan C, Ward T. Correction factor for continuous monitoring of wood smoke fine particulate matter. *Aerosol Air Qual Res* 2011;11:315–22. <http://dx.doi.org/10.4209/aaqr.2010.08.0072>.
- [35] Diapoulis E, Chaloulakou A, Spyrellis N. Indoor and outdoor PM concentrations at a residential environment, in the Athens area. *Glob NEST J* 2008;10(2):201–8.
- [36] EN12341:1998. Determination of the PM<sub>10</sub> fraction of suspended particulate matter – reference method and field test procedure to demonstrate reference equivalence of measurement methods.
- [37] Freitas MC, Reis MA, Marques AP, Almeida SM, Farinha MM, Oliveira O, et al. Monitoring of environmental contaminants: 10 years of application of k0-INAA. *J Radioanal Nucl Chem* 2003;257(3):621–5. <http://dx.doi.org/10.1023/A:1025405003508>.
- [38] Freitas MC, Almeida SM, Reis MA, Ventura MG. Neutron activation analysis: still a reference method for air particulate matter. *J Radioanal Nucl Chem* 2004;262:235–9. <http://dx.doi.org/10.1023/B:JRNC.0000040880.76829.0d>.
- [39] Almeida SM, Freitas MC, Reis M, Pinheiro T, Felix PM, Pio CA. Fifteen years of nuclear techniques application to suspended particulate matter studies. *J Radioanal Nucl Chem* 2013;297:347–56. <http://dx.doi.org/10.1007/s10967-012-2354-1>.
- [40] Almeida SM, Silva AV, Freitas MC, Marques AM, Ramos CA, Silva AI, et al. Characterization of dust material emitted during harbour activities by k0-INAA and PIXE. *J Radioanal Nucl Chem* 2012;291:77–82. <http://dx.doi.org/10.1007/s10967-011-1279-4>.
- [41] Almeida SM, Ramos CA, Marques AM, Silva AV, Freitas MC, Farinha MM, et al. Use of INAA and PIXE for multipollutant air quality assessment and management. *J Radioanal Nucl Chem* 2012;294:343–7. <http://dx.doi.org/10.1007/s10967-011-1473-4>.
- [42] Dung HM, Freitas MC, Blaauw M, Almeida SM, Dionísio I, Canha NH. Quality control and performance evaluation of k0-based neutron activation analysis at the Portuguese research reactor. *Nucl Instrum Methods A* 2010;622:392–8. <http://dx.doi.org/10.1016/j.nima.2010.04.003>.
- [43] Almeida SM, Almeida-Silva M, Galinha C, Ramos CA, Lage J, Canha N, et al. Assessment of the Portuguese k0-INAA laboratory performance by evaluating internal quality control data. *J Radioanal Nucl Chem* 2014;300:581–7. <http://dx.doi.org/10.1007/s10967-014-2987-3>.
- [44] Almeida-Silva M, Almeida SM, Gomes JF, Albuquerque PC, Wolterbeek HT. Determination of airborne nanoparticles in elderly care centers. *J Toxicol Environ Health A* 2014;77(14–16):867–78. <http://dx.doi.org/10.1080/15287394.2014.910157>.
- [45] ICRP. *Annals of the ICRP* 24:240. International Commission on Radiological Protection. 3rd ed. New York: Macmillan; 1994.
- [46] Portaria n.º 353-A/2013. Ministérios do Ambiente, Ordenamento do Território e Energia, da Saúde e da Solidariedade, Emprego e Segurança Social.
- [47] Weschler. Ozone in indoor environments: concentration and chemistry. *Indoor Air* 2000;10:269–88. <http://dx.doi.org/10.1034/j.1600-0668.2000.010004269.x>.
- [48] Wang LZ, Emmerich SJ, Persily AK, Lin CC. Carbon monoxide generation, dispersion and exposure from indoor operation of gasoline-powered electric generators under actual weather conditions. *Built Environ* 2012;56:283–90. <http://dx.doi.org/10.1016/j.buildenv.2012.03.016>.
- [49] Lee SC, Sanches L, Fai HK. Characterization of VOCs, ozone, and PM<sub>10</sub> emissions from office equipment in an environmental chamber. *Built Environ* 2001;36:837–42. [http://dx.doi.org/10.1016/S0360-1323\(01\)00009-9](http://dx.doi.org/10.1016/S0360-1323(01)00009-9).
- [50] Destaillets H, Maddalena RL, Singer BC, Hodgson AT, McKone TE. Indoor pollutants emitted by office equipment: a review of reported data and information needs. *Atmos Environ* 2008;42(7):1371–88.
- [51] Pegas PN, Alves CA, Evtyugina MG, Nunes T, Cerqueira M, Franchi M, et al. Indoor air quality in elementary schools of Lisbon in Spring. *Environ Geochem Health* 2011;33:455–68. <http://dx.doi.org/10.1007/s10653-010-9345-3>.
- [52] EPA. An introduction to indoor air quality (IAQ) – volatile organic compounds (VOCs). Obtained in Fevereiro de 2012. de United States Environmental Protection Agency; 2011. <http://www.epa.gov/iaq/voc2.html>.
- [53] Hoskins J. Health effects due to indoor air pollution. *Indoor Built Environ* 2003;12:427–33. <http://dx.doi.org/10.1177/1420326X03037109>.
- [54] Shin SH, Jo WK. Longitudinal variations in indoor VOC concentrations after moving into new apartments and indoor source characterization. *Environ Sci Pollut Res* 2013;20:3696–707. <http://dx.doi.org/10.1007/s11356-012-1296-z>.
- [55] Hanninen O. Combining CO<sub>2</sub> data from ventilation phases improves estimation of air exchange rates. In: *Healthy buildings 2012*: Austrália; 2012.
- [56] Canha N, Almeida S, Freitas M, Taubel M, Hanninen O. Winter ventilation rates at primary schools: comparison between Portugal and Finland. *J Toxicol Environ Health A* 2013;76(6):400–8. <http://dx.doi.org/10.1080/15287394.2013.765372>.
- [57] Dimitroulopoulou C. Ventilation in European dwellings: a review. *Built Environ* 2012;2012(47):109–25. <http://dx.doi.org/10.1177/1420326X13481786>.
- [58] Persily A. Evaluating building IAQ and ventilation with indoor carbon dioxide. *ASHRAE Trans* 1997;103(2).
- [59] Satish U, Mendell MJ, Shekhar K, Hotchi T, Sullivan D, Streufert S, et al. Is CO<sub>2</sub> an indoor pollutant? Direct effects of low-to-moderate CO<sub>2</sub> concentrations on human decision-making performance. *Environ Health Perspect* 2012;120(12):1671–7. <http://dx.doi.org/10.1289/ehp.1104789>.
- [60] Zhou Y, Levy J. The impact of urban street canyons on population exposure to traffic-related primary pollutants. *Atmos Environ* 2008;42:3087–98. <http://dx.doi.org/10.1016/j.atmosenv.2007.12.037>.
- [61] Kao L, Nanagas K. Carbon monoxide poisoning. *Med Clin North Am* 2005;89:1161–94.
- [62] Corsi RL, Siegel JA, Chiang C. Particle resuspension during the use of vacuum cleaners on residential carpet. *J Occup Environ Hyg* 2008;5(4):232–8. <http://dx.doi.org/10.1080/15459620801901165>.

- [63] Canha N, Ribeiro M, Freitas MC, Almeida M, Almeida SM, Cabo S, et al. Fungi, bacteria and pollens seasonally quantified at 3 basic schools in Lisbon. In: ASHRAE IAQ 2010 conference: airborne infection control – ventilation, IAQ & energy, Kuala Lumpur, Malaysia, 10–12 November. ISBN 978-1-936504-04-6.
- [64] Almeida SM, Freitas MC, Pio CA. Neutron activation analysis for identification of African mineral dust transport. *J Radioanal Nucl Chem* 2008;276:161–5. <http://dx.doi.org/10.1007/s10967-007-0426-4>.
- [65] Almeida SM, Freitas MC, Repolho C, Dionísio I, Dung HM, Caseiro A, et al. Characterizing air particulate matter composition and sources in Lisbon. *Port J Radioanal Nucl Chem* 2009;281(2):215–8. <http://dx.doi.org/10.1007/s10967-009-0113-8>.
- [66] Almeida SM, Silva AI, Freitas MC, Dung HM, Caseiro A, Pio CA. Impact of maritime air mass trajectories on the Western European Coast urban aerosol. *J Toxicol Environ Health A* 2013b;76(4–5):252–62. <http://dx.doi.org/10.1080/15287394.2013.757201>.
- [67] Freitas SM, Farinha MM, Pacheco AMG, Ventura MG, Almeida SM, Reis MA. Atmospheric selenium in an industrialized area of Portugal. *J Radioanal Nucl Chem* 2005;263(3):711–9. <http://dx.doi.org/10.1007/s10967-005-0647-3>.
- [68] Mason B, Moore CB. *Principles of geochemistry*. New Jersey: John Wiley; 1982.
- [69] Farinha MM, Freitas MC, Almeida SM. Air quality control monitoring at an urban and industrialized area. *J Radioanal Nucl Chem* 2004;259:203–7. <http://dx.doi.org/10.1023/B:JRNC.0000017288.21685.85>.
- [70] Buonanno G, Marini S, Morawska L, Fuoco FC. Individual dose and exposure of Italian children to ultrafine particles. *Sci Total Environ* 2012;438:271–7. <http://dx.doi.org/10.1016/j.scitotenv.2012.08.074>.
- [71] Gomes J, Bordalo JCM, Albuquerque PC. Monitoring exposure to airborne ultrafine particles in Lisbon, Portugal. *Inhal Toxicol* 2012;24(7):425–33. <http://dx.doi.org/10.3109/08958378.2012.684077>.
- [72] Ntziachristos L, Polidori A, Phuleria H, Geller MD, Sioutas C. Application of a diffusion charger for the measurement of particle surface concentration in different environments. *Aerosol Sci Technol* 2007;41:571–80.